



Visible Light Activated Fe-N-SiO₂/TiO₂ Photocatalyst: Providing an Opportunity for Enhanced Photocatalytic Degradation of Antibiotic Oxytetracycline in Aqueous Solution

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ABSTRACT

To probe advantages in Fe-N-SiO₂/TiO₂ nanocomposite system, the visible photocatalytic degradation of the nonbiodegradable antibiotic oxytetracycline (OTC) by unsupported TiO₂ and its modified composites by incorporating each of the Fe, N, and SiO₂ dopants under a series of conditions were investigated. The structural and optical properties as well as the morphology of the prepared nanocomposites were also characterized applying Fourier transform infrared (FT-IR), X-ray diffraction, photoluminescence spectroscopy, UV-visible diffuse reflectance spectra and field emission scanning electron microscopy/Energy-dispersive X-ray spectroscopy (FESEM/EDX). In order to develop two models portraying appropriate functional relationships between two main responses (OTC removal efficiency and its specific removal rate (SRR)) and four numerical variables (OTC concentration, catalysis loading, initial pH and reaction time), two separate multivariate analysis pathways under response surface methodology (RSM) were taken. The results obtained all came down to the maximum SRR (220 OTC mg_{OTC removed}/g_{cat.} h) found at the maximum catalyst dosage of 1.5 g/l, and acidic pH of 3 after 0.5 h. Furthermore, the Fe-N-SiO₂/TiO₂ proved a stable photocatalytic activity during three subsequent reusability experiments, shedding light on its reliable potential for future application.

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1. INTRODUCTION

As a member of one of the most commonly worldwide used classes of antibiotic chemicals aiming significantly for disease prevention and its treatment, oxytetracyclines (OTC) has been frequently detected in the environment, mostly involving water bodies [1, 2]. The environmental contamination caused by OTC is usually originated from pharmaceutical manufacture, livestock farming, and agricultural discharges [3, 4].

Due to be extremely resistant to biodegradation [5], a large number of studies focused on alternative non-biological processes for OTC removal [6]. Providing the opportunity of being carried out at room temperature and

atmospheric pressure while obtaining high levels of mineralization of various pollutants to mainly water, CO₂, and inorganic compounds, throughout the recent decades, photocatalytic technology as an advanced oxidation process (AOP) has been showing great promise in antibiotics degradation [7]. Performing under visible light irradiation is the key to the practical application of photocatalysts. Meanwhile, TiO₂ mediated photocatalysis has been highlighted as one of the most promising approaches to degrade stable organic contaminants in water by producing hydroxyl radicals [8, 9]. Given their sufficiently optical/electric properties, non-toxic nature, high photocatalytic capacity, and long-term chemical stability, metal oxide semi-conductors

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including TiO₂ have been proven in recent publications as potential photocatalysts for successful oxidation as well as antibacterial inactivation of OTC [10, 11].

Possessing a large band gap (3.2 eV) limiting its activation to only ultraviolet (UV) region and low quantum efficiency caused by high e⁻/h⁺ recombination, TiO₂ meets a major drawback in its wide application [12]. Many efforts have been devoted to overcome the limitations improving the visible driven photoactivity and ultimate photocatalytic efficiency of titania photocatalysts [13]. In order to obtain TiO₂-based visible light activated photocatalysts, many strategies including metal ions doping [14], nonmetal ions doping [15], and coupling with other semiconductors containing narrower band gaps [16] have been adopted. Despite some controversial efficiency results obtained under different experimental conditions, introduction of transition metal cations including ferric ion (Fe³⁺) in titania has yet been considered an effective approach towards enhancing the photocatalytic properties within the visible light region [17-19]. Being able to be replaced into TiO₂ network structure due to its electron configuration, different optimum amounts of Fe³⁺ incorporated in TiO₂ as: 0.3 wt.% [20], 0.5 wt.% [21], and 1.0 wt.% [22] have been reported in the literature.

As the leading element which has been investigated extensively in the literature due to its favorable features as a dopant, nitrogen (N) can be easily introduced in the crystal lattice of TiO₂ while suppressing the recombination rate of the photogenerated electron/hole pairs and ultimately facilitating its visible-light photocatalytic activity [23]. Park et al. [24] investigated the enhanced visible driven photocatalytic degradation of methylene orange (MO) N-doped TiO₂ nanocomposite fabricated by using graft polymerization. In another research, Bergamonti and colleagues [25] reported the synthesis of N-TiO₂ photocatalysts employing a variety of precursors as oxysulfate (N-TiA), tetraisopropoxide (N-TiU), and titanium (IV) for MO and rhodamine B (RhB) photodegradation. Based on the results obtained, N-TiA led to the highest photocatalytic activity owing to its capacity absorb higher wavelengths of light in the UV-visible spectrum. Meanwhile, suggesting greater surface area, higher capacities for adsorbing pollutant molecules, and higher photoactivity while being bond-conjugated to TiO₂, SiO₂ has also become a potential candidate in order to obtain enhanced photodegradation efficiencies in the last decades [26]. Chun and coworkers [27-29] thoroughly investigated the adsorption and photodegradation of various dyes using the SiO₂/TiO₂ photocatalyst, shedding light on its high potential.

The present study attempts to compare visible driven photocatalytic degradation of OTC by unsupported TiO₂ nanoparticles and with TiO₂ nanocomposites modified by doping a metal (Fe³⁺), nonmetal (N), and coupling an oxide semi-conductor compound (SiO₂) in order to have

a better understanding of the advantages these modifications can bring along. The experiments successfully resulted in the fabrication of the modified Fe-N-SiO₂/TiO₂ photocatalyst. Following the physicochemical and optical characterization of the as-obtained photocatalysts, As the main objective of the research, effects of four factors including catalyst loading (0.5-1.5 g/l), OTC concentration (50- 200 mg/l), initial pH (3-11), and irradiation time (0.5-6.5 h) were thoroughly investigated.

2. MATERIAL AND METHODS

2.1. Materials Oxytetracycline (OTC) pure powder ≥ 99.9% was ordered from Sigma, USA. Tetra n-butylorthotitanate (TNBOT) ≥ 98%, Tetraethyl Ortho Silicate (TEOS) ≥ 98%, Ethanol 96%, HCl 37%, NH₃ as the nitrogen source, and Fe(NO₃)₃.9H₂O as Fe⁺³ source were all supplied from Merck, Germany. All the other reagents used were all analytical grade without further requirement for purification. All aqueous solutions were prepared with ultra-pure water.

2.2. Synthesis OF Fe-N-SiO₂/TiO₂ Catalysis Fe-N-SiO₂/TiO₂ photocatalyst was synthesized via the sol-gel technique [21]. 25 ml of TNBOT was dissolved in 100 ml of ethanol under magnetic stirring for 30 min (A solution). Then, 1.6 ml of ethanol was added to TEOS under stirring for 10 min followed by the addition of 4 ml of HCl 37% + 3 ml of ultra-pure water and then further stirred for 30 min (B solution). Next, C solution was prepared by dissolving 0.148 g of Fe(NO₃)₃.9H₂O in 20 ml of ethanol under stirring for 10 min. Ultimately, the mixture of NH₃, and B and C solutions were added to A solution followed by being stirred for about 6 h. The final solution was aged for 48 h. The prepared sample was then dried at 100 °C for approximately 10 h, ultimately being calcinated at 450 °C for 2.5 h.

2.3. Characterization of the Prepared Photocatalysts Using KBr pellets containing the powder samples, an FT-IR spectrometer (MAGNA-560) was applied to record Fourier Transform Infra-Red (FT-IR) spectra. X-ray diffractometer (a Rigaku D-max C III) with Ni-filtered K_α radiation and Philips XL30 microscope with accelerating voltage of 10 kV made it possible to obtain X-ray diffraction (XRD) patterns and scanning electron microscopy (SEM) images, respectively leading to a better observation and estimation of the photocatalysts morphology and their structural properties. UV-Vis spectrophotometer (Rayleigh UV 2601 model) and the photoluminescence spectrometer using a photomultiplier tube (Perkin Elmer LS55) were used to study DRS and PL analysis, respectively.

2. 4. Photodegradation Experiments

Photodegradation performance of the synthesized photocatalysts was evaluated by degrading the synthetic OTC wastewater. Reaction suspensions were prepared by adding appropriate amounts of the as-prepared photocatalyst powders into 200 ml of the synthetic OTC wastewater. Solutions were then irradiated by ultraviolet (UV) and visible (Vis) lights for specific time intervals. Experiments were performed for 390 min, and liquid aliquots (5 ml) were withdrawn every 60 min. The samples were then centrifuged to remove the existing particles prior to being analyzed by UV-Vis spectrophotometry. COD concentration of the samples were measured using standard method (5220D) procedures [30]. In each sample, OTC concentration was measured according to its absorbance detected by a UV-Vis spectrophotometer (DR 5000, Jenway) at maximum wavelength of 348 nm. In order to better compare the results, bare Titania was also tested along with the other samples under the same experimental conditions.

2. 5. Experimental Set Up As shown in Figure 1, the photodegradation tests in the laboratory scale were conducted in a suspended batch photoreactor. Being located in a black box, the photoreactor consisted of a 200 ml Pyrex-glass cell. Irradiation sources were a 100-watt tungsten lamp and a UV lamp (HITACHI, emission: 365 nm, constant intensity 60 mW/cm^2) which were located at the upper part of the photoreactor. The vertical distance between the solution and the irradiation source was 15 cm. Furthermore, a magnetic stirrer was applied to keep the solution uniformed by providing constant agitation throughout the experiments at the constant temperature of $25 \pm 1 \text{ }^\circ\text{C}$.

2. 6. Experimental Design As one of the most common experimental methods, response surface methodology (RSM) under Design Expert software (Stat-Ease Inc., version 11.1.2.0) was employed to design the photodegradation tests as well as their mathematical modeling leading to further data analysis [31]. Accordingly, four independent numerical factors including initial OTC concentration, catalyst concentration, reaction time, and initial pH were assessed in terms of their effects on the OTC removal process performance and modeled at 5 levels ($\alpha = \pm 0.5$) by inscribed central composite design (CCID) as shown in Table 1. According to the following equation, a total of 30 operating conditions [32] (Equation (1)) were designed.

$$N = 2^{k-q} + 2 \times k + n_c \quad (1)$$

where k , q , and n_c are considered as the number of factors, a fraction of the number factor (which is zero for full factorial design) and the replicate number of the central point, respectively [32]. The values of the

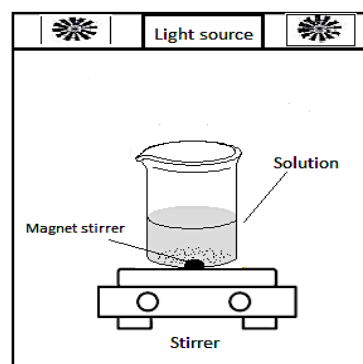


Figure 1. The schematic of the employed photocatalytic batch reactor

TABLE 1. Experimental range and levels of the independent variables

Variables	Range and levels				
	-1	- α	0	+ α	1
A- Initial OTC concentration, mg/l	50	87.5	125	162.5	200
B-Catalyst concentration, g/l	0.5	0.75	1	1.25	1.5
C- Initial pH	3	5	7	9	11
D- Reaction time, h	0.5	2	3.5	5	6.5

responses obtained in the experiments are presented in Table 2.

3. RESULTS AND DISCUSSION

3. 1. Photocatalytic Activity Experiments: Optimization of Fe-N-SiO₂/TiO₂ Each prepared photocatalyst was subjected to a photocatalytic reaction for OTC removal at certain conditions (OTC concentration = 50 ppm and photocatalyst concentration = 0.5g/l) under visible light for 6.5 h (390 min). The prepared samples included pure TiO₂ and its combination with different mass ratios of the dopants (Fe-TiO₂, N-TiO₂, SiO₂-TiO₂, and Fe-N-SiO₂-TiO₂). As observed in Figure 2, pure TiO₂ did not exhibit much visible photoactivity (only 40%). The optimum weight fractions of Fe, N, and SiO₂ as the dopants were obtained at 0.05, 5, and 5 wt. %, respectively, and were further applied in the modified photocatalysts.

Undergoing the same visible light irradiation for 390 min, all of the doped nanocomposites led to higher visible light activities in comparison to the pure TiO₂ as shown in Figure 2. Fe-N-SiO₂/TiO₂, resulting in 91 % OTC removal efficiency, showed the maximum visible-light photocatalytic activity compared to the other doped photocatalysts: Fe-TiO₂ (80%) N-TiO₂ (70%), SiO₂-TiO₂ (64%).

TABLE 2. Experimental conditions for the photocatalytic process

Run	Factor 1	Factor 2	Factor 3	Factor 4	Response 1	Response 2
	A: OTC concentration	B: Catalyst concentration	C: Initial pH	D: Reaction time	OTC removal eff.	SRR
	mg/l	g/l		h	%	mg OTC removed/ g cat. h
1	50	0.5	3	6.5	80.26	12.34
2	50	0.5	3	0.5	56.36	78.35
3	50	0.5	11	0.5	25.26	31.86
4	50	0.5	11	6.5	45.93	7.06
5	50	1.5	3	6.5	86.93	4.45
6	50	1.5	11	0.5	22.09	8.1
7	50	1.5	11	6.5	62.26	3.19
8	50	1.5	3	0.5	60.26	24.2
9	87.5	1	7	3.5	92.1	20.5
10	125	0.75	7	3.5	71.09	29.75
11	125	1	7	3.5	73.35	23.1
12	125	1	7	3.5	75.9	25.28
13	125	1	7	5	83.84	19.01
14	125	1	9	3.5	75	23.561
15	125	1	7	3.5	75.1	25.21
16	125	1	7	3.5	73.4	25.99
17	125	1	7	2	70.9	35.45
18	125	1	7	3.5	73.1	23.2
19	125	1	7	3.5	73.9	25.8
20	125	1	5	3.5	84.5	26.5
21	125	1.25	7	3.5	83.3	20.1
22	162.5	1	7	3.5	78.2	32.04
23	200	0.5	3	0.5	33.5	173
24	200	0.5	11	6.5	31.56	19.42
25	200	0.5	11	0.5	23	119.73
26	200	0.5	3	6.5	53.56	32.96
27	200	1.5	3	6.5	83.98	51.68
28	200	1.5	11	6.5	42.9	8.8
29	200	1.5	3	0.5	70.45	347.2
30	200	1.5	11	0.5	28.3	51.7

The improved visible driven photocatalytic activity of Fe-doped TiO₂ (80%) is attributed to the narrower band gap leading to the enhanced generation of the charge carriers and higher photocatalytic activity [17, 33, 34]. Furthermore, acting as shallow traps in the titania lattice, Fe³⁺ cations lead to a reduction in electron-hole recombination properties, hence increasing not only the electron-hole pair lifetime but also the possibility of reactions between the created electron-hole pair. As for N-doped TiO₂, owing to the formation of a mid-gap (N^{2p}) with an intermediate energy level over top of the

(O^{2p}) valence band [35], introducing nitrogen into the TiO₂ lattice results in its narrower band gap and ultimately enhanced its photocatalytic activity under visible light irradiation. While in SiO₂-TiO₂ composite, the semi-conductor cations entering the lattice of TiO₂ not only provides more available surface for photocatalytic activity, but also enhances its surface acidity. Consequently, coupling SiO₂ with TiO₂ leads to preferentially adsorbing more hydroxyl groups on the surface of the composite and therefore, decreasing its contamination by OTC [36].

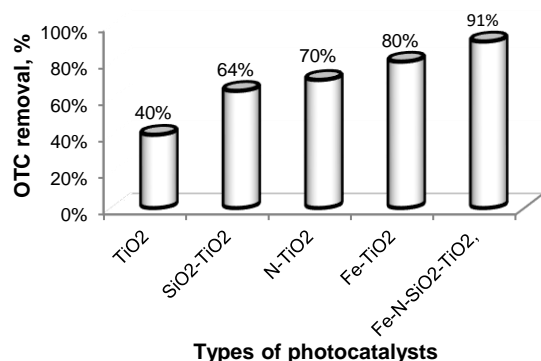


Figure 2. Visible-driven OTC removal efficiency by different photocatalysts at the OTC concentration of 50 ppm and photocatalyst concentration of 0.5g/l for 390 min

Relying on the aforementioned reasons supporting the improved visible driven photocatalytic activity in three modified nanocomposites, Fe-N-SiO₂/TiO₂ obtained the highest OTC removal efficiency, simultaneously benefiting from the addition of Fe³⁺, N, and SiO₂.

3. 2. Characterization of Fe-N-SiO₂/TiO₂ Photocatalyst

3. 2. 1. FT-IR Spectroscopy Figure 3 illustrates the FT-IR spectra of the prepared samples. The bands at 400-800 are related to the vibration of Ti-O and Ti-O-Ti bonds in all of the samples [30]. The presence of dissolved or atmospheric CO₂ in the samples also resulted in a peak at 2400 cm⁻¹. Furthermore, the weak transmittance bands at around 1630 and 3100-3600 cm⁻¹ can be assigned respectively to the bending and stretching vibrations of hydroxyl groups of the adsorbed water molecules in the samples [37]. Owing to the nitrogen atoms being substituted into the TiO₂ network in N-TiO₂ and Fe-N-SiO₂-TiO₂, the peak at 1440 cm⁻¹ was obtained in Figure 3 [38].

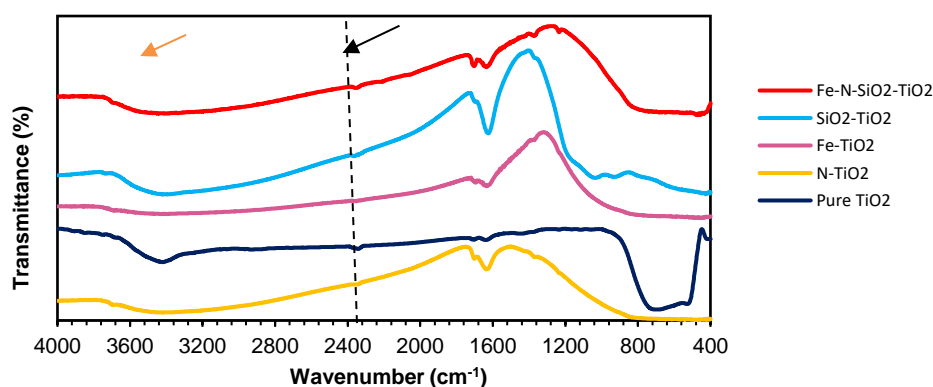


Figure 3. FT-IR spectra of the prepared samples

The asymmetric vibration of Si-O-Si bonds for SiO₂-TiO₂ composites was observed at approximately 1060 cm⁻¹ [50]. The stretching vibrations of Si-OH, SiO groups and Si-O-Ti bonds can be implied by the absorption band at about 900~1000 cm⁻¹ [39].

3. 2. 2. XRD Patterns The XRD pattern was employed to identify crystal structure of the prepared pure TiO₂, SiO₂-TiO₂, N-TiO₂, Fe-TiO₂ and Fe-N-SiO₂-TiO₂ photocatalysts as illustrated in Figure 4. At all samples, the peaks are noted to tetragonal anatase crystalline phase of TiO₂. The diffraction peaks of the 2θ values at 25.1°, 37.25°, 38.63°, 48.12°, 54.48°, 62.54°, 68.9°, 70.57°, and 75.15° were observed which are related to the crystalline phases of anatase. The XRD pattern for the doping amount of Fe (0.05 wt. %) was too low to be detected by XRD [40]. The crystalline size (D) of Fe-N-SiO₂/TiO₂ nanoparticles was calculated by using Debye Scherrer's formula ($D = 0.9\lambda / \beta \cos\theta$), where λ is the wavelength of the X-rays and β is the full width at the half maximum intensity (FWHM) [41]. The results of the calculations for the XRD data showed that the range of particle sizes of Fe-N-SiO₂/TiO₂ NPs are between 19.64 to 46.78 nm, which is close to and supports SEM results.

3. 2. 3. SEM Analysis The SEM images of the Fe-N-SiO₂/TiO₂ nanoparticles are depicted in Figures 5(a)-5(c). As observed in Figure 5(a), the sizes of the nanoparticles were in the range of 22.14-50.87 nm. As an extra provision to calculate the size of the catalyst, Figure 5b presented the particle size distribution of Fe-N-SiO₂/TiO₂ nanoparticles in the range of 10 to 45 nm, in which the most common size range included the values from 26 to 30 nm [42]. The SEM results are the clear evidence that the uniformly distributed nanoparticles are spherical in shape. Providing more active areas for photocatalytic degradation, the porous structure of the as-prepared photocatalysts has also been proven by these images [40]. The EDX result in Figure 5(c) confirms the presence of Fe, N, Si, Ti and O elements in the nanocomposite.

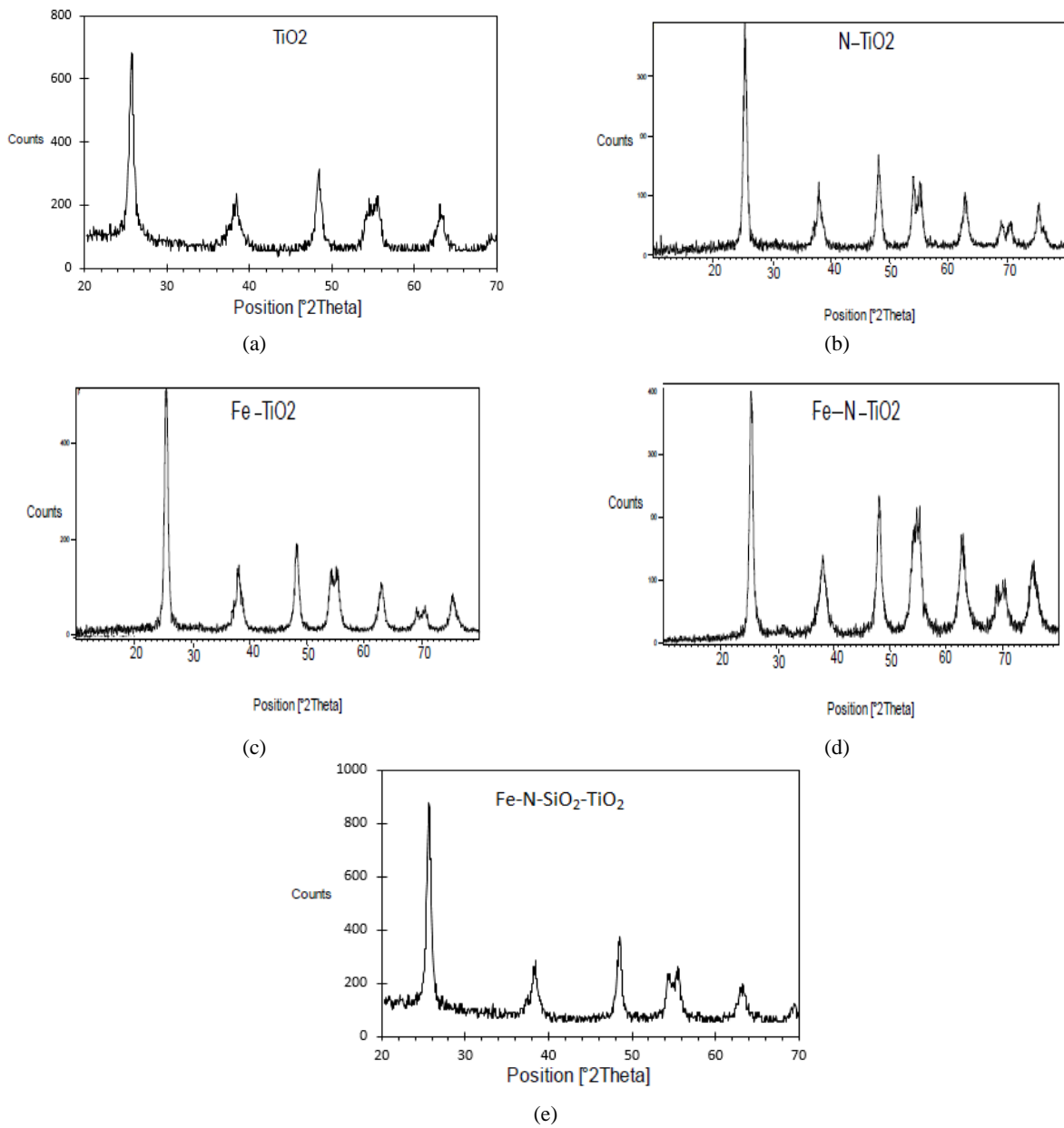
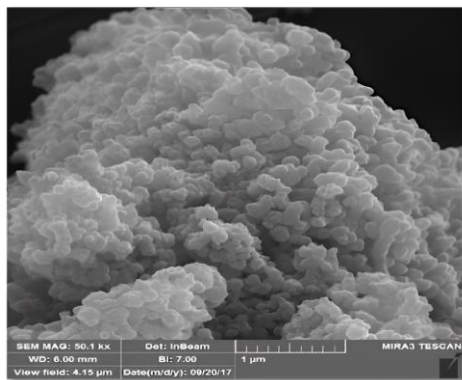
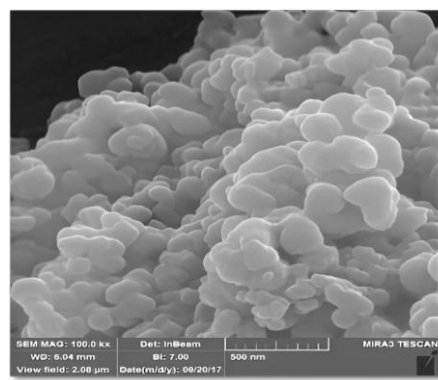


Figure 4. XRD patterns of the prepared samples



(a)



(b)

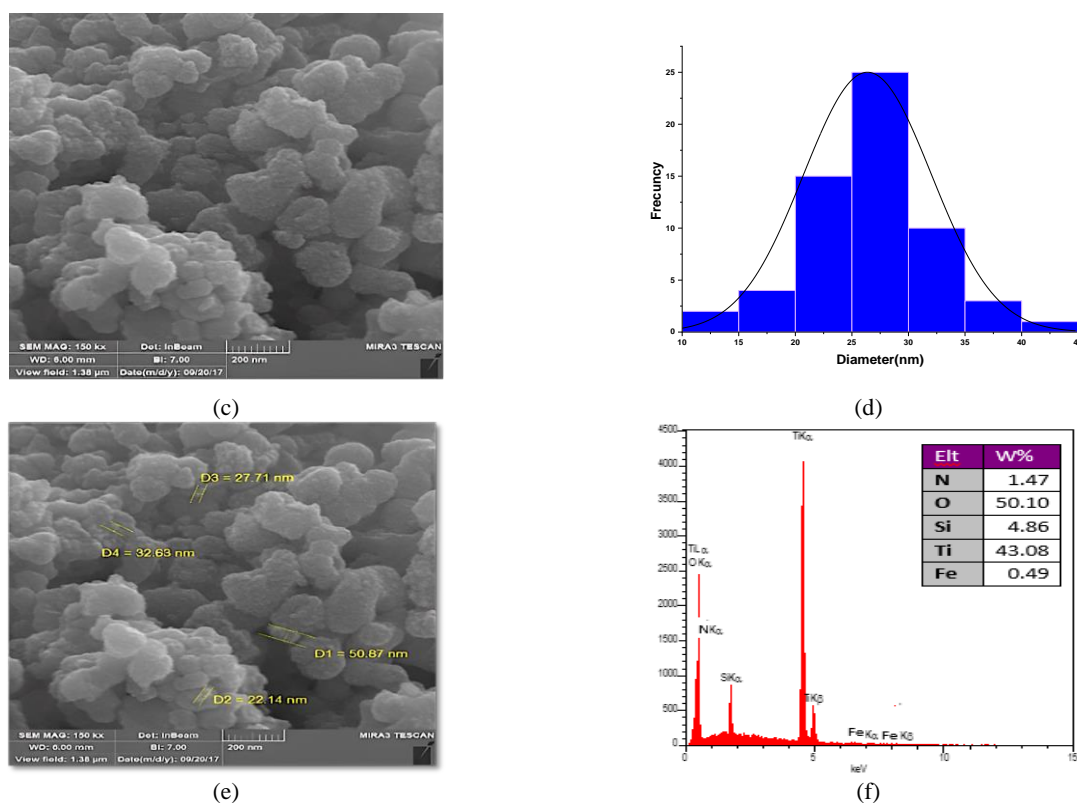


Figure 5. SEM images and EDX spectrum of Fe-N-SiO₂/TiO₂

3. 2. 4. Optical Properties The optical absorption of the synthesized photocatalyst was assessed by UV-Vis DRS analysis and its result was indicated in Figure 6(a). The strong absorption in the visible light region implies the vital role of dopant agents (Fe and N elements) in the reduction of band gap and extension of the absorption edge into a visible light range. Figure 6(b) shows the PL emission spectrum of pure TiO₂ and Fe-N-SiO₂/TiO₂ photocatalysts. The PL intensity of modified nanocomposite is lower than pure TiO₂ due to incorporating Fe, N, and SiO₂ on the TiO₂ surface [43, 44]. Both oxygen vacancies and lattice defects in the structure of the modified TiO₂ sample can act as charge carrier trapping centers and infer the recombination rate of photogenerated h⁺/e⁻ pairs decline.

3. 3. Photodegradation Process Analysis And Modeling

In order to fit the CCID experimental results obtained in OTC photodegradation runs, a response surface model using standard analysis of variance (ANOVA) was developed and the subsequent results are provided in Table 3. Two models, reduced quadratic and reduced 2FI, containing the confidence level of 99 % (probability value <0.0001) were selected for the OTC removal efficiency and specific removal rate (SRR), respectively.

Implying the degree of significance and influence of the coefficients on the responses, the probability values (p-values) and F-values are given in Tables 3 and 4. The insignificant model terms with significantly higher than 0.05 p-values need to be removed from the model equation simplifying the model. As a result, according to p-values for OTC removal, the significant model terms include A (OTC concentration), B (catalyst concentration), C (initial pH), D (reaction time), AB, AD, BC, and B². Meanwhile, for SRR, A, B, C, D, AB, AC, and BC are considered significant model terms. Given the F-values and also multilateral contributions of each variable in the final equation, significant model terms were indicated in terms of their order and subsequent influences on both responses.

Evaluating the accuracy and variability of the model for the OTC removal efficiency and SRR, the high values of the coefficient of determination ($R^2 = 0.9487$ and $R^2 = 0.9570$, respectively) ensures a satisfactory adjustment of the developed models to the experimental data [45]. Furthermore, being very close to the corresponding R^2 value, the adjusted R^2 values of respectively 0.9291 and 0.9458 further confirm the adaptability of the aforementioned models. Predicted R^2 values of both responses are in good agreement with experimental values (Predicted R^2 was about 0.8721 and 0.8900 for OTC removal and SRR, respectively).

According to Table 3, the adequate precision values for OTC removal efficiency and SRR were 23.8605 and 44.2257, respectively, further proving that the models are desirable. Figures 7(a) and 7(b) depict the measured data versus the predicted ones for both models. Furthermore, coefficient of variation (C.V. %) is an important concept allowing us to predict variables within and outside data sets [45]. High reliability and accuracy of experimental work in this study could also be explained by C.V. values of 9.07% and 6.99% for OTC removal and SRR, respectively.

Favoring more effective contacts between the OTC and hydroxyl radicals, the conceivable effect of longer reaction time on the response was observed in both plots.

The three-dimensional (3D) response surface plots represented in Figures. 8 and 9 are applied to assess the interaction effects of the aforementioned variables on both responses. The concentration of OTC was held constant at an average level (125 mg/l) in all of the plots given in Figure 8. In the meantime, increasing the OTC concentration to the maximum level might have an inhibiting role due to the saturation of the photocatalyst active sites with OTC molecules. As a result, the poisoning phenomenon might occur at high OTC loadings, resulting in less production of hydroxyl radicals and superoxide radical anions thereby weakening the degradation efficiency [46].

TABLE 3. ANOVA results for the selected models regarding OTC removal efficiency.

Source	Sum of Squares	df	Mean Square	F-value	p-value			
Model	12684.80	8	1585.60	48.53	< 0.0001	Significant	Final Equation in Terms of Coded Factors	
A-OTC conc.	378.72	1	378.72	11.59	0.0027			
B-Cat. conc.	785.50	1	785.50	24.04	< 0.0001			
C-Initial pH	3750.09	1	3750.09	114.78	< 0.0001	OTC removal eff. =		
D-Reaction time	1848.22	1	1848.22	56.57	< 0.0001	+78.18		
AB	227.10	1	227.10	6.95	0.0154	-4.79		A
AD	186.73	1	186.73	5.72	0.0263	+6.90		B
BC	144.84	1	144.84	4.43	0.0475	-15.08		C
B ²	5363.59	1	5363.59	164.17	< 0.0001	+10.58		D
Residual	686.09	21	32.67			+3.77		AB
Lack of Fit	679.76	16	42.48	33.57	0.0005	Significant	-3.42	AD
Pure Error	6.33	5	1.27			-3.01	BC	
Cor Total	13370.89	29				-27.58	B ²	

Fit Statistics						
Std. Dev	Mean	C.V. %	R ²	Adjusted R ²	Predicted R ²	Adeq Precision
5.72	63.01	9.07	0.9487	0.9291	0.8721	23.8605

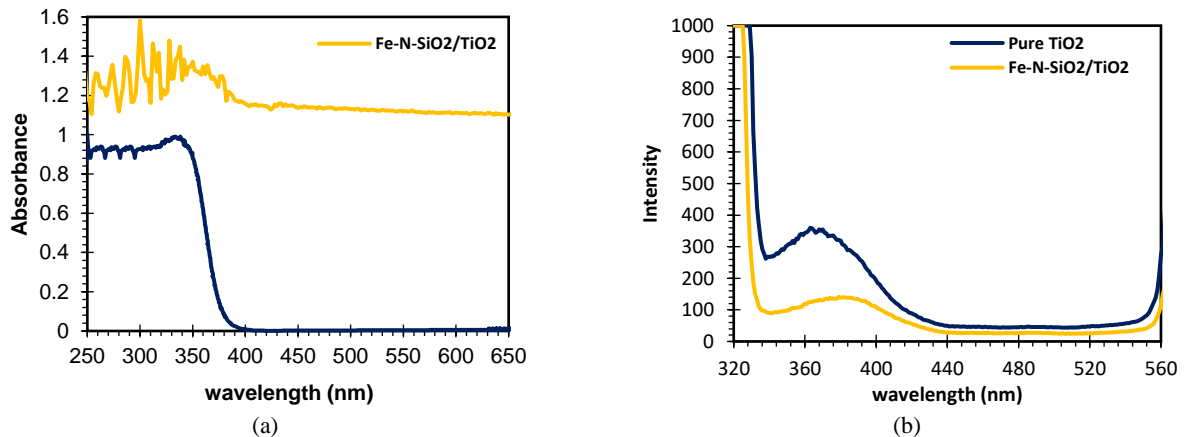


Figure 6. (a) DRS and (b) PL spectra of the prepared photocatalysts

Figures 8(a) and 8(b) demonstrate the influence of B (catalyst concentration) and C (initial pH) on OTC removal efficiency. Being highly influenced by the initial pH of the solution in terms of its photolytic and hydrolytic stability in deionized water while presenting pKa values of 3.22, 7.46, and 8.94, OTC shows repulsive or attractive interactions with the amphoteric nature of the TiO₂-based photocatalyst in different pH levels. Thus, pH is expected to play a key role in OTC photocatalytic degradation experiments. As such, Figures 8(a) and 8(b) show that at an average OTC concentration of 125 mg/l, as initial pH level goes from acidic to basic (3 to 11), OTC removal efficiency decreases, proving the detrimental effect of basic pH levels to OTC mineralization. The observed decreasing trend can be

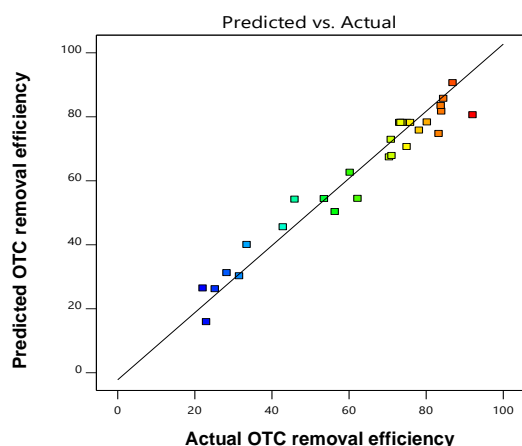
attributed to the presence of negatively charged OTC molecules at high pH (9 and above), with a high electrical density on the ring system, which tend to attract increasing concentration of reactive species such as an hydroxyl radicals, thus affecting OTC mineralization [47]. The low degree of OTC mineralization implies that consumed OTC mainly transforms to byproducts.

However, it is noted that progressive increase in the catalyst loading from 1 to 1.5 g/l led to a decreasing trend in the response. As observed, at higher than 1g/l catalyst loadings, turbidity increases leading to the unfavorable light scattering phenomenon, and thus, the reduction of light penetration into the suspension [48]. Consequently, the decreased photoactivated volume of the suspension resulted in lower OTC removal efficiencies.

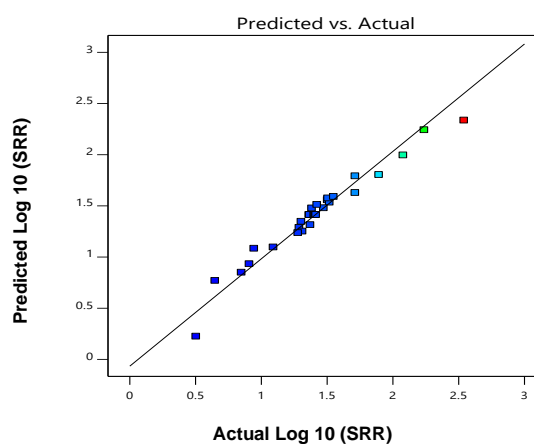
TABLE 4. ANOVA results for the selected models regarding SRR

Source	Sum of Squares	df	Mean Square	F-value	p-value		
Model	5.04	7	0.7194	82.82	< 0.0001	Significant	
A-OTC conc.	1.73	1	1.73	199.64	< 0.0001	Final Equation in Terms of Coded Factors	
B-Cat. conc.	0.2903	1	0.2903	33.42	< 0.0001		
C-Initial pH	0.6458	1	0.6458	74.34	< 0.0001		
D-Reaction time	2.07	1	2.07	237.85	< 0.0001		
AB	0.1769	1	0.1769	20.36	0.0002		
AC	0.0335	1	0.0335	3.86	0.0623		
BC	0.0889	1	0.0889	10.24	0.0041		
Residual	0.1911	22	0.0087			$\text{Log}_{10}(\text{SRR}) = +1.41 + 0.3242A - 0.1326B - 0.1978C - 0.3539D + 0.1051AB - 0.0746BC$	
Lack of Fit	0.1885	17	0.0111	21.23	0.0016		Significant
Pure Error	0.0026	5	0.0005				
Cor Total	5.23	29					

Fit Statistics						
Std. Dev	Mean	C.V. %	R ²	Adjusted R ²	Predicted R ²	Adeq Precision
0.0988	1.41	6.99	0.9570	0.9458	0.8900	44.2257



(a)



(b)

Figure 7. Predicted vs. actual values plot for (a) OTC removal efficiency and (b) SRR

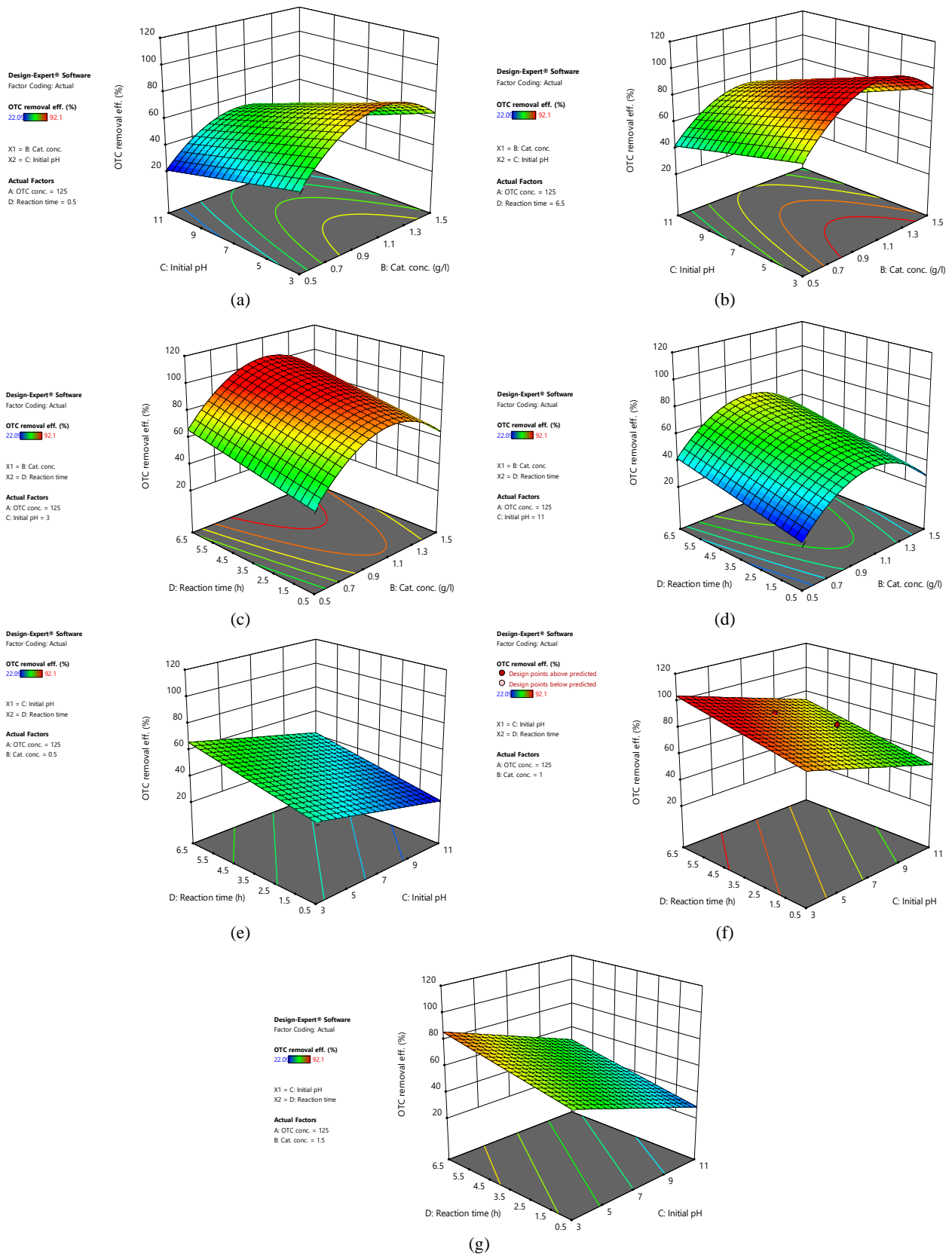


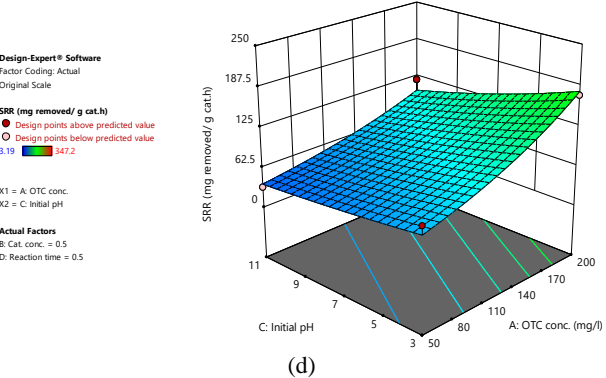
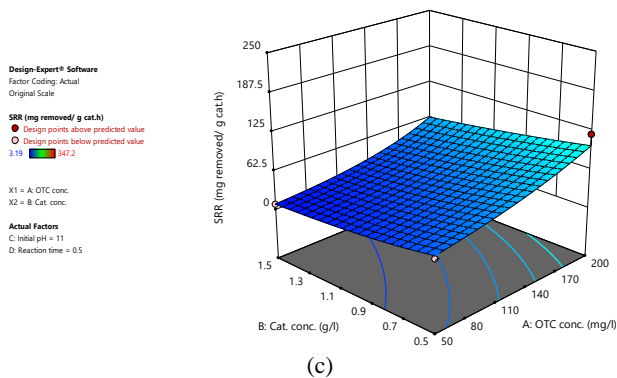
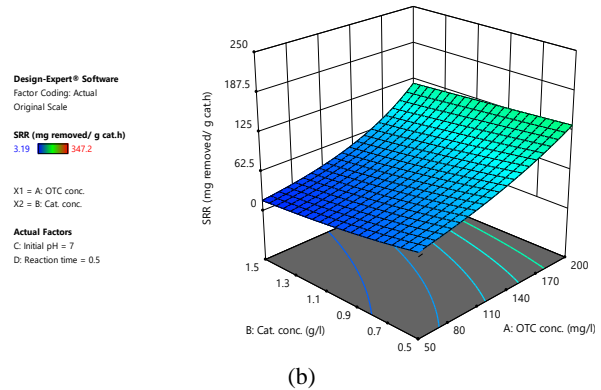
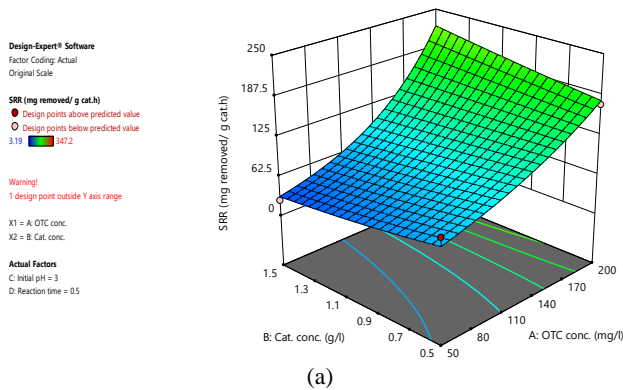
Figure 8. 3D surface plots of OTC removal for the most important pairs of factors: a and b) catalyst concentration and initial pH, c and d) catalyst concentration and reaction time, e, f and g) initial pH and reaction time

As indicated from both plots, the reaction time presented a linear positive effect on the response from 0.5 to 6.5 h. The influence of B (catalyst concentration) and D (reaction time) on the response is plotted in Figures 8(c) and 8(d). The aforementioned impact of catalyst loading on the response is represented again in these two plots, further proving the two consecutive ascending and descending trends observed below and above 1 g/l of catalyst loadings, respectively.

Favoring more effective contacts between the OTC and hydroxyl radicals, the conceivable effect of longer reaction time on the response was observed in both plots. Similarly, initial pH, showed a linear increasing impact on the response at the constant concentration of 125 mg/l of OTC.

Figures 8(e), 8(f), and 8(g) display the 3D surface plots of the response as a function of C (initial pH) and D (reaction time) at three different levels of catalyst concentration (0.5, 1, and 1.5 g/l) and average OTC concentration (125 mg/l). These plots further confirm the linear positive impacts of both initial pH and reaction time on the OTC removal efficiency. Meanwhile, the catalyst loading depicted a parabolic trend opening downward. It means that it has an increasing impact on removal efficiency until reaching catalyst concentration of 1 g/l and then reversing its impact from 1 to 1.5 g/l. In order to assess the overall photocatalytic performance, treatment capacity, and mass efficiency of the prepared

photocatalyst, specific OTC removal rate (SRR) was plotted as a function of different variables in Figures 9(a)-9g. The reaction time remained at the minimum level of 0.5 h. It is clear that reaction time has a decreasing impact on the SRR, referring to the possible surface poisoning of the catalyst. Figure 8 (a, b, and c) represents the response surface plots for the SRR as a function of A (OTC concentration) and B (catalyst concentration) at the minimum reaction time (0.5 h) and different initial pH levels (3, 7, and 11). As observed in these plots, initial pH has a decreasing impact on SRR owing to the repulsive interaction between the catalyst surface and contaminant at higher pH levels. Furthermore, Figures 9d and 9(e) represent the effects of A (OTC concentration) and C (initial pH) at constant minimum reaction time (0.5 h) and minimum and maximum levels of catalyst loading, further highlighting the impacts of catalyst loading and initial pH on this response. The 3D plots given in Figures 9(f) and 9(g) indicate the effects of B (catalyst concentration) and C (initial pH) at the minimum reaction time (0.5 h) and minimum and maximum OTC concentrations (50 and 200 mg/l, respectively). It is noted that at higher OTC concentrations and acidic pH of 3, catalyst loading shows an increasing trend, further confirming the enhanced mass efficiency of the photocatalyst in this condition. According to the plots shown in Figures 9(a), 9(e), and 9(g), maximum SRR of



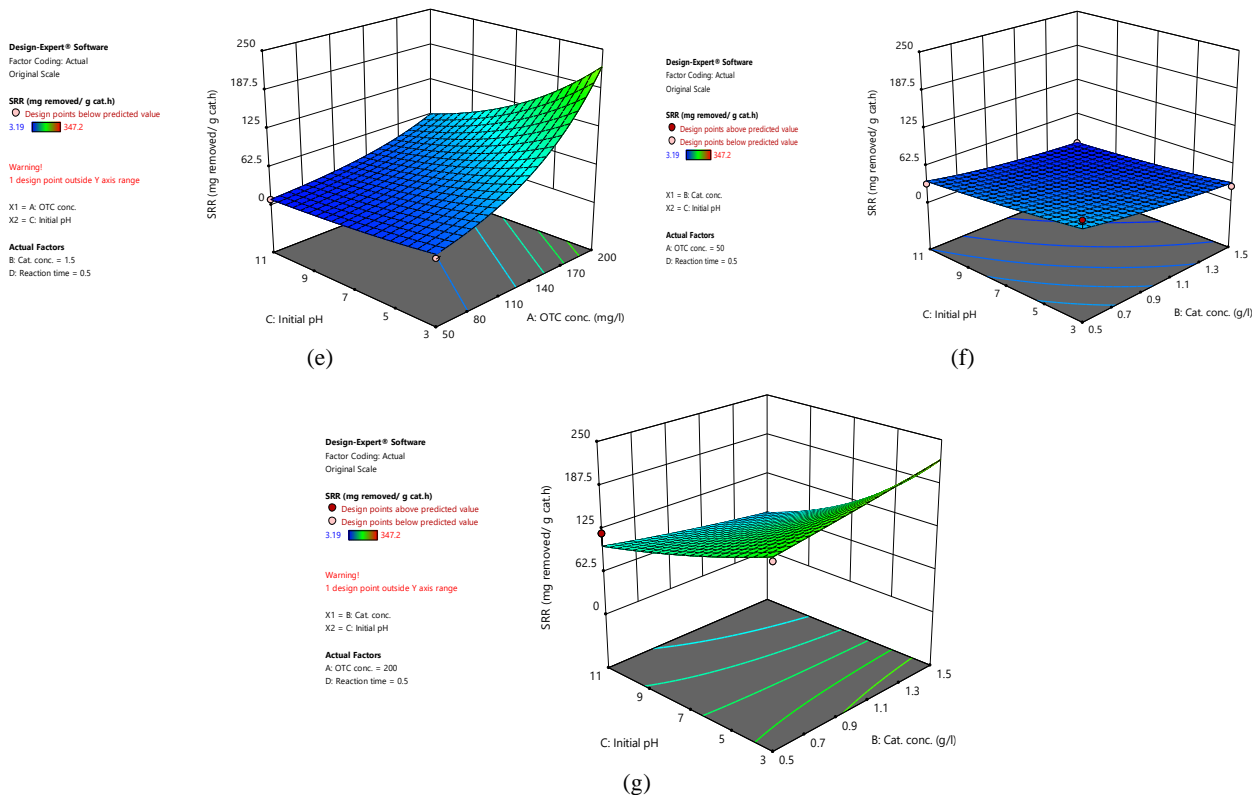


Figure 9. 3D surface plots of OTC removal for the most important pairs of factors: a, b, and c) catalyst concentration and OTC concentration, d and e) OTC concentration and initial pH, f and g) catalyst concentration and initial pH

approximately 220 mg OTC removed/g cat. h was obtained at initial pH of 3, minimum reaction time of 0.5 h, maximum catalyst loading of 1.5 g/l, and maximum OTC concentration of 200 mg/l.

3. 4. Photolytic Degradation of OTC under Controlled pH

A visible photolytic degradation experiment was carried out at the variation of pH values to monitor the OTC removal and mineralization without the addition of the photocatalyst. The results obtained after 6.5 h are given in Figure 9. As can be seen in Figure 10, after 390 min of visible light irradiation, only the range containing a minimum level of 3% to a maximum level of 22% of the initial OTC concentration was degraded at pH levels of 3 to 11 (3, 4.4, 5, 7, 9, and 11), respectively. As aforementioned in the previous part (Figures 8(a) and 8(b) analysis), as pH increases to 9 and above, the negatively charged OTC molecules tend to facilitate the photolytic degradation of OTC and, thus, the formation of more recalcitrant intermediates to mineralization, proving that photocatalysis plays a smaller role at such higher pH levels [47].

Overall, the negligible degree of OTC photolytic mineralization is enough proof of merely sufficient degradation of OTC molecules into more stable

intermediates in comparison to their photocatalytic degradation [47]. Furthermore, in order to assess the contribution of dark adsorption in OTC removal efficiency, the same experiments were carried out in darkness resulting in a minimum 1% to maximum 17% at pH values of 3 to 11 respectively.

3. 5. Reusability of the Fe-N-SiO₂-TiO₂ Photocatalyst

As shown in Figure 11, the reusability of the prepared modified photocatalyst used in the reactions was investigated by the degradation of OTC after 3 cycles under the same conditions. The used photocatalyst after each cycle was regenerated with 60-min aeration under visible light. As can be seen from the results, the photocatalyst could be reused without a significant loss in its activity after 3 successive runs and only an approximate reduction of 2-4% in the process performance was observed. The aforementioned loss might be caused by gradual poisoning and insignificant catalyst loss during supernatant removal after each cycle. Overall, the reusability results indicated the sufficient stability and reliability of the Fe-N-SiO₂/TiO₂ nanocomposite proving its potential for practical application.

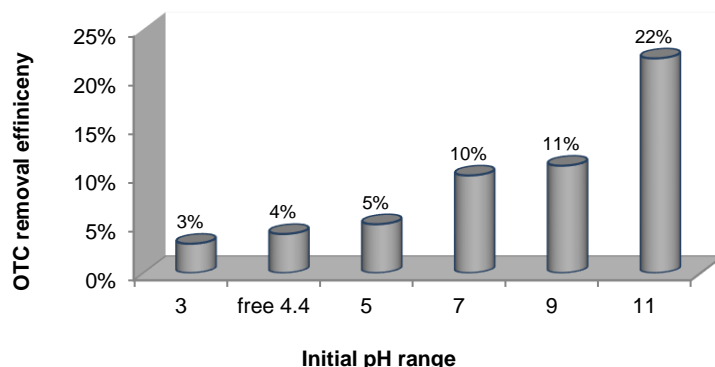


Figure 10. Photolytic degradation of OTC at different pH values (3-11) in the absence of Fe-N-SiO₂/TiO₂ under visible light irradiation for 390 min

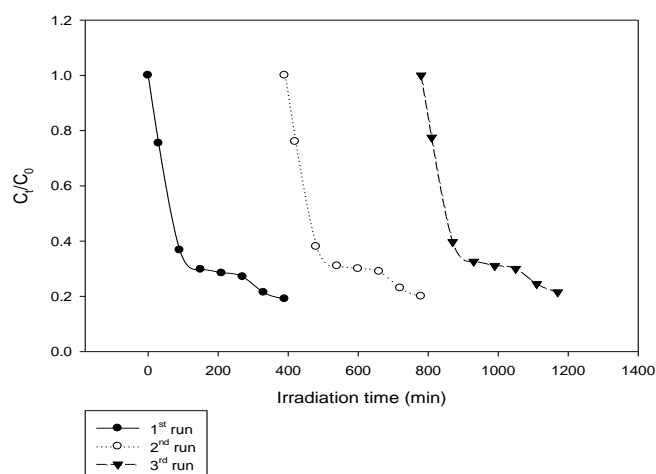


Figure 11. Reusability test of the Fe-N-SiO₂-TiO₂ photocatalyst degrading OTC during three successive runs after regular periodic regeneration using 60-min aeration under visible light

4. CONCLUSION

In the present study, in order to obtain the enhanced photocatalytic decomposition of the nonbiodegradable OTC compounds under visible light, the Fe-N-SiO₂/TiO₂ photocatalyst was successfully synthesized by sol-gel technique and its physical properties were characterized via FT-IR and SEM. The effects of OTC concentration, catalyst concentration, initial pH, and reaction time on the OTC removal efficiency and its SRR were evaluated using CCID under RSM. The reaction time had a positive effect on the OTC removal efficiency while a reverse impact was observed for the initial pH. As for the assessment of the optimum overall performance of the prepared photocatalyst, maximum SRR of 220 mg OTC removed/g cat. h at certain conditions (OTC Conc. = 200 mg/l, catalyst conc. = 1.5 g/l, pH = 3, and reaction time = 0.5 h) was obtained. Furthermore, the Fe-N-SiO₂/TiO₂

nanocomposite could easily be recovered and reused after 3 cycles without considerable loss in its photocatalytic activity.

5. ACKNOWLEDGEMENT

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Persian Abstract

چکیده

برای بررسی مزایای سیستم نانوکامپوزیت Fe-N-SiO₂/TiO₂، تخریب فوتوکاتالیستی آنتی بیوتیک اکسی تتراسایکلین (OTC) توسط TiO₂ و کامپوزیت های اصلاح شده آن با ترکیب هر یک از دوپنت ها شامل N، Fe و SiO₂ تحت شرایط خاصی مورد بررسی قرار گرفتند. خواص ساختاری و نوری و همچنین مورفولوژی نانوکامپوزیت های تهیه شده نیز با استفاده از طیف (FT-IR)، پراش اشعه ایکس، طیف سنجی فوتولومینسانس، طیف های DRS و FESEM/EDX مشخص شد. به منظور توسعه دو مدل که روابط عملکردی مناسب بین دو پاسخ اصلی (بازده حذف OTC و سرعت حذف ویژه آن (SRR)) و چهار متغیر عددی (غلظت OTC، غلظت کاتالیست، pH اولیه و زمان واکنش) را نشان می دهند، دو مسیر تجزیه و تحلیل چند متغیره مجزا تحت روش سطح پاسخ (RSM) بکار گرفته شد. نتایج به دست آمده همگی به حداکثر (220 SRR (mg OTC removed/g cat. h) در حداکثر دوز کاتالیزور ۱.۵ گرم در لیتر و pH اسیدی ۳ پس از ۰.۵ ساعت کاهش یافت. علاوه بر این، Fe-N-SiO₂/TiO₂ یک فعالیت فوتوکاتالیستی پایدار را در طی سه آزمایش پشت سر هم ثابت کرد و پتانسیل قابل اعتماد آن را برای کاربردهای آینده روشن کرد.
